

Geographical patterns and partition of turnover and richness components of beta-diversity in faunas from Tocantins river valley

Tonial, MLS.^a, Silva, HLR.^{a,b}, Tonial, IJ.^a, Costa, MC.^a,
Silva Júnior, NJ.^{a,b} and Diniz-Filho, JAF.^{c*}

^aSystema Naturae Consultoria Ambiental Ltda, Rua 58, 217, Quadra Bloco 4, Lote16, Jardim Goiás, CEP 74810-250, Goiânia, GO, Brazil

^bDepartamento de Biologia, Pontifícia Universidade Católica de Goiás – PUC Goiás, Goiânia, GO, Brazil

^cDepartamento de Ecologia, Instituto de Ciências Biológicas – ICB, Universidade Federal de Goiás – UFG, Goiânia, GO, Brazil

*e-mail: diniz@icb.ufg.br

Received July 6, 2011 – Accepted October 19, 2011 – Distributed August 31, 2012
(With 1 figure)

Abstract

There has been a resurging interest in patterns of β -diversity, especially by the mechanisms driving broad-scale, continental and global patterns, and how partitioning β -diversity into richness (or nestedness) and turnover components can be linked with such mechanisms. Here we compared two recent methodologies to find richness and turnover components of β -diversity, using a large regional scale dataset of mammal, bird, reptiles and amphibian species found in seven regions of Central, North and Northeastern Brazil. As well as a simple comparison of the metrics available, we analyzed spatial patterns (i.e., distance-decay similarity) and the effects of biome type in these components using raw and partial Mantel tests. Our analyses revealed that turnover estimated using Baselga's (2010) approach is slightly higher than the estimate using Carvalho's et al. (2012) approach, but all analyses show consistent spatial patterns in species turnover using both methods. Spatial patterns in β -diversity revealed by Mantel tests are also consistent with expectations based on differential dispersal abilities. Our results also reinforce that spatial patterns in β -diversity, mainly in the turnover components expressing faunal differentiation, are determined by a mix of broad scale environmental effects and short distance spatially-structured dispersal.

Keywords: faunal similarity, β -diversity, turnover, richness, nestedness.

Padrões geográficos e desdobramento entre componentes de substituição e riqueza da diversidade-beta em faunas do Centro e do Nordeste do Brasil

Resumo

Tem havido um interesse crescente nos padrões de diversidade- β , particularmente acerca dos mecanismos subjacentes a esses padrões em escalas continentais e globais, e sobre o modo como esses mecanismos podem ser associados ao desdobramento desses padrões em componentes de riqueza e substituição (*turnover*) de espécies. Neste artigo, nós comparamos duas metodologias desenvolvidas recentemente para realizar esse desdobramento, utilizando conjuntos regionais de espécies de mamíferos, aves, répteis e anfíbios em sete regiões no Nordeste, Norte e no Centro do Brasil. Além de comparar os dois conjuntos de índices utilizados para o desdobramento dos componentes da diversidade- β , foram analisados os padrões espaciais e os efeitos do bioma no qual são encontradas as regiões nesses componentes por meio de testes de Mantel e testes de Mantel parciais. Embora não tenham sido encontradas diferenças na magnitude das estimativas de diversidade- β entre os grupos, foi possível detectar diferenças significativas entre os padrões espaciais, principalmente no componente de substituição, sugerindo que os padrões na variação de composição são determinados por um balanço entre efeitos ambientais em grandes escalas e por processos de dispersão espacialmente estruturados.

Palavras-chave: comparação de faunas, diversidade- β , riqueza de espécies, substituição.

1. Introdução

Since the classical paper of Whittaker (1972), the partition of diversity into multiple components, such as α , β and γ , has been done using several techniques. Conceptually, α -diversity is the number of species in a given sample or locality, whereas the β -diversity indicates the difference among localities (the γ being, thus, the sum of α and β component, providing an estimate of regional diversity) (see Tuomisto, 2010a, b; Anderson et al., 2011 for a recent review and redefinition of these concepts). In the last 10 years, some papers started to investigate these patterns using continental and global datasets (e.g., Gaston et al., 2007; Qian, 2009; Melo et al., 2009) and these triggered a whole new set of questions regarding the mechanisms driving species turnover at broad geographical scales.

It has been recognized for a while that the several metrics available for estimating β -diversity capture different components of differences among sites or regions (e.g., Harrison et al., 1992; Lennon et al., 2001; Koleff et al., 2003), but Baselga (2010) recently proposed that these components could be formally partitioned using these different metrics. The idea is that the difference among species in two faunas could be due to a replacement of species (turnover) or due to a shift in species richness (or nestedness component, in which a poorer-fauna is composed by a subset of species of a richer fauna – see Almeida-Neto et al. 2012). Moreover, Baselga (2010) focused on the interpretation of the balance among nestedness and turnover components and on how this balance could be linked to macroecological patterns and processes associated to the more general latitudinal gradients in biodiversity. He proposed a geographical pattern in the nestedness component, which would be stronger at higher latitude. This pattern was confirmed by Dobrovolski et al. (2012), based on beta-diversity variation in New World mammals, birds and amphibians (see also Melo et al., 2009). Baselga's (2010) partition has been widely used to describe patterns in beta diversity and to support hypotheses about ecological and evolutionary processes driving biodiversity patterns at distinct scales (e.g., Hortal et al., 2011; Andrew et al., 2012).

Despite the large interest in Baselga's (2010) approach to partition beta-diversity into distinct components, the details of the metrics used by him have been criticized on different grounds (e.g., Almeida-Neto et al., 2012; Carvalho et al., 2012), and some improvements have been proposed based on the same reasoning (Carvalho et al., 2012; but see Baselga 2012). Here we analyzed a data set of faunal similarity of terrestrial vertebrates in Central, North and Northeast Brazil to answer the following questions: 1. What is the difference between Baselga's (2010) and the newer approach proposed by Carvalho et al. (2012) for partition of β -diversity in a real, rather than simulated or artificial, dataset? 2. Do organisms with restricted-dispersal, such as amphibians, possess stronger β -diversity than dispersal-prone organisms, like mammals and birds? and 3. Do the different components of β -diversity (i.e., turnover and richness) possess distinct spatial patterns and

do these spatial patterns vary among groups of vertebrates with distinct dispersal?

2. Methods

2.1. Data

We obtained data of presence-absence of mammals (186 species), birds (578 species), squamate reptiles (214 species) and amphibians (80 species) for seven regions in Central, North and Northeastern Brazil. Data (i.e., species lists) were compiled from several sources and by intensive sampling (see below) in these regions (Eletronorte, 1985a, b; Systema Naturae Consultoria Ambiental Ltda, 1996, 1999, 2008, 2010a, b, 2011; Passamani, 2004; Pavan and Dixo, 2004; Pinheiro, 2004; Puerto and Barbarini, 2004; Silva Junior and Sites, 1995). Data for these regions were originally derived from studies conducted to access faunal characteristics in broad scale dam buildings found in areas from the Cerrado biome (including Serra da Mesa, GO; Cana Brava, GO; São Salvador, GO/TO and; Peixe Angical, TO), in the Amazon biome (Tucuruí, PA) and in transitions between Cerrado and Amazonia (Luís Eduardo Magalhães, TO and Estreito, MA/TO).

Data for compiling faunal lists came from faunal inventories and rescues, as well as monitoring programs conducted for several years in the regions nearby the dam impact area. Full data matrix used for the analyses is available from the authors upon request. Despite a lack of homogeneity due to necessary variation in sampling efforts across habitats within localities and among different teams conducting faunal analyses in these regions, these studies represent fauna variation at regional scales. The potential problems in the data certainly would affect more detailed analysis of faunal patterns at a finer geographical scale, but are suitable for comparing different regions because they tend to represent exhaustive species lists for each region. We expect thus that our ecological conclusions are unlikely to be qualitatively affected by data issues.

2.2. Partitioning β -diversity into richness and turnover components

We analyzed the dataset described above using the approaches proposed by Baselga (2010) and Carvalho et al. (2012). The details of computation and derivation of the formulae used can be found in these original papers, and only a brief description of the formulae used for calculations is given below.

All metrics used for β -diversity among two localities or regions, say 1 and 2, start by the well-known computation of counts of a (number of species common to both regions or localities), b (number of species exclusive of region or locality 1), c (number of species exclusive of region or locality 2) and d (species not found in both regions or localities). According to Baselga's (2010) approach, the overall β -diversity is computed by Sorensen's dissimilarity index, given by (Equation 1)

$$\beta_{\text{SOR}} = (b + c)/2a + b + c \quad (1)$$

which can be additively partitioned into turnover (β_{SIM}) and nestedness (β_{NEST}) components as (Equation 2)

$$\beta_{SOR} = \beta_{SIM} + \beta_{NEST} \quad (2)$$

where (Equation 3)

$$\beta_{SIM} = \min(b, c)/a + \min(b, c) \quad (3)$$

and (Equation 4)

$$\beta_{NEST} = [(b - c)/(a + b + c)] * [a/a + \min(b, c)] \quad (4)$$

Notice that these quantities and indices are expressed here as dissimilarities to better match the notion of diversity (i.e., larger numbers expressing more diversity), and not as similarities among faunas, as more frequently done in ecological studies. Therefore, β_{SIM} express turnover between the two regions or localities as the proportion of the species-poorer region, whereas β_{NEST} is better viewed as a proportion of richness difference (the β_{SOR} , which is the first term in the equation of β_{NEST}), multiplied by the proportion of the species-poorer region or locality that is nested in the species-richer region (the second term in the equation). Almeida-Neto (2012), among other things, criticized the use of the term “nestedness” in this context mainly because other more specific measurements have been proposed and widely tested in a different context (e.g., Almeida-Neto et al., 2008).

Carvalho et al. (2012) criticized the metrics used by Baselga’s (2010) and pointed out that β_{SIM} overestimates the turnover component, and proposed different metrics for making the partition. The overall reasoning is the same, but the overall beta-diversity is computed based on Jaccard’s complement, rather than Sorensen, and is given by (Equation 5)

$$\beta_{CC} = [(b + c)/(a + b + c)] \quad (5)$$

where partition is given by (Equation 6)

$$\beta_{CC} = \beta_{-3} + \beta_{rich} \quad (6)$$

Where β_{-3} and β_{rich} are the metrics for differences in species turnover and richness (“nestedness”, in Baselga’s, (2010) terminology), respectively, and are given by (Equation 7)

$$\beta_{-3} = 2 * [(\min(b, c) / (a + b + c))] \quad (7)$$

and (Equation 8)

$$\beta_{rich} = [(b - c)/(a + b + c)] \quad (8)$$

We computed the metrics described above for our dataset for total β -diversity (i.e., all 1058 species), and its richness and turnover components, as well as for the four groups of terrestrial vertebrates separately. We computed pairwise matrices for all dissimilarities among the seven regions and obtained the mean values of all pairwise comparisons. The relative importance of turnover, according to Baselga (2010) and Carvalho et al. (2011) metrics, was thus obtained by (Equation 9)

$$T_B (\%) = (\beta_{SIM} / \beta_{SOR}) * 100 \quad (9)$$

and (Equation 10)

$$T_C (\%) = (\beta_{-3} / \beta_{CC}) * 100 \quad (10)$$

2.3. Spatial patterns in β -diversity components

The spatial patterns in β -diversity and its components, calculated using the formulae described above, were analyzed by a Mantel test (Manly, 1998) between the pairwise matrices of β -diversity and the geographic distance among sites, transformed to natural logarithms to linearize the relationship. The matrix correlation of the Mantel test was tested using 999 random permutations.

The seven regions were located in different biomes, as pointed out above (4 in Cerrado, 1 in the Amazonia and 2 in Cerrado-Amazonia transitions), and this can affect the spatial patterns of β -diversity and its components. We added this biome effect as a discrete model matrix (see Manly, 1985; Legendre and Legendre, 1998) by performing a Mantel test between pairwise matrices of β -diversity and its components with a model matrix, with 1 indicating that the pair of regions compared is in the same biome and zero elsewhere. We also analyzed patterns in beta diversity by partial Mantel tests, so it is possible to test the effect of biome independently of geographical distance and of the geographical distance independently of the biome.

β -diversity indices were computed using a Q-BASIC program written for this paper, whereas Mantel tests were performed using SAM 4.0 (Rangel et al., 2006, 2010).

3. Results

There is not a large difference in the overall β -diversity among the four groups of vertebrates, for both Baselga’s (2010) and Carvalho et al. (2012) approaches (Table 1). β_{SOR} tends to be smaller than β_{CC} , but their relative magnitude of difference among the groups are quite similar. For Baselga’s (2010) methodology, higher β -diversity is found in mammals, whereas when using Carvalho’s et al. (2012) methodology the higher β -diversity is observed for amphibians. Using both methodologies, reptiles are the group with smaller β -diversity. The proportion of turnover also did not show a strongly consistent pattern in both methodologies, except for much larger values for reptiles. For all groups, however, the turnover component is higher for Baselga’s (2010) methodology, coherent with Carvalho’s et al. (2012) conclusion that β_{SIM} tends to give highest values for this component.

Spatial patterns in β -diversity components, analyzed by the Mantel test, show consistent patterns among groups of organisms and components (Table 1). Focusing initially on the overall β -diversity, it is possible to see that matrix correlations with geographical distances are higher for amphibians and reptiles ($r > 0.7$ for all cases, and going up to $r = 0.84$ for reptiles) than for mammals and birds ($r < 0.65$ in all cases), for both methodologies used. Moreover, these spatial patterns in overall β -diversity were consistently determined by its turnover component, and richness or nestedness component only showed significant spatial pattern for amphibians, regardless of the relative balance

between richness/turnover components for all groups in the distinct methodologies.

However, the spatial component is also correlated with differences in the biomes in which each region is found (Mantel's $r = 0.861$), so it is necessary to evaluate the effect of biomes on β -diversity, as well as the partial effects of geographical distances and biome. It is possible to see that the effect of the biome is usually higher than the geographical distances and they remain significant when geographical distances are kept constant in partial Mantel tests (Table 1). On the other hand, effects of geographical distance keeping biomes are usually not significant, although the rank of patterns among groups described above are also maintained. Notice, however, that because the effect of the biome is evaluated using a discrete model matrix there is less power to detect effects due to a small number of regions within each biome class.

4. Discussion

Our analyses of faunal variation among the seven regions studied in Central Brazil reveal consistent patterns of β -diversity, and can be well interpreted both in the methodological and ecological sense, providing clear answers to all three questions formulated.

We observed some differences between Baselga's (2010) and Carvalho et al. (2012) methodologies for partition of β -diversity into its richness and turnover components, although the overall patterns for all groups studied are consistent among organisms. As pointed out by Carvalho et al. (2012), the turnover component of the partition seems to be slightly larger when using β_{SIM} , even after taking into account the differences between the two metrics for overall β -diversity (i.e., the Sorensen dissimilarity used by Baselga's (2010) is smaller than Jaccard used by Carvalho et al. (2012), even so the turnover proportion is higher in all cases). Thus, Carvalho's et al. (2012) formulae for the partition avoids confounding the idea of nestedness and differences in β -diversity due to differences in richness (see Almeida-Neto, 2012), but patterns in partial regressions are clearer using Baselga's (2010) original formulae.

The differences of β -diversity among groups are not clearly defined in our analyses, and cannot be directly linked to potential dispersal or life-history patterns. One could argue that the partition of species into such large groups such as mammals, birds, reptiles and amphibians are not accurate to represent even broad patterns of dispersal (i.e., especially because mammals and birds, for example, mix species with very large and very small body sizes and with variable life histories, which directly affects dispersal). Although this is certainly true, Dobrovolski et al. (2012) found clear differences among mammals, birds and reptiles in their proportion of turnover on a continental scale (i.e., with amphibians presenting the smaller amount of turnover, as expected by its lower potential dispersal ability). The more homogeneous results among groups found here, thus may be better viewed as a consequence of the relatively

small scale of the study and of the variations in sampling efforts among groups (i.e., Dobrovolsky's et al., 2012 was based on the extent of occurrence data which is not potentially disturbed by sampling effects).

On the other hand, much more interesting and consistent results were found when dealing with spatial patterns and effects of biome in β -diversity. Baselga (2010) pointed out and proposed that, on macroecological scales, one would expect to find a higher proportion of nestedness components at a higher latitude, mainly because they are more clearly affected by historical events of climate change, such as glaciations. This was supported by Dobrovolski et al. (2012) analyses, which also showed that the magnitude of this effect interacts with dispersal ability, as pointed out before (see also Qian, 2009). This idea is also consistent with the mechanisms usually associated with species richness latitudinal gradients, in which decreasing ability to persist in colder environments is usually linked with past adaptations and ancestral conditions (the niche conservatism; Wiens and Donoghue, 2004; Hawkins et al., 2006; Rangel et al., 2007), geographically-structured patterns in speciation/extinction balance (see Mittelbach et al., 2007) or environmental tolerance (Rangel and Diniz-Filho 2005). Thus, in short, environmental components acting throughout evolutionary time tend to reduce richness at a higher latitude, so differences among localities in different latitudes tend to be better explained by the "nestedness" component.

However, at finer spatial scales, a distinct pattern is observed, especially in the tropics, because no strong environmental gradients are observed driving richness patterns at regional scales. Consequently, a relatively high turnover component is expected and, more importantly, spatial patterns will appear in the turnover of species more clearly than in the richness or nestedness component. For instance, a distance-decay of β -diversity is expected under Hubbell's (2001) neutral theory (see Cottenie, 2005; Soininen et al., 2007; but see Diniz-Filho et al., 2012 for a more complex evaluation) and also this spatial structure should appear more clearly for groups with more restricted dispersal ability, exactly as found here. At the same time, different biogeographic regions, appearing due to both ecological and historical events, are also expected to generate more differences in turnover and species composition than in richness (e.g., Vasconcelos et al., 2011). Indeed, our analyses show both effects, although biome effects are much stronger than pure geographical distances and remain significant even in partial Mantel tests. On smaller scales, although effects are not significant, they remain comparatively structured among groups, suggesting that within biomes the β -diversity among the regions studied possess a strong stochastic component and are mainly determined by spatially-structured dispersal. Analyses in more refined scales, using more local datasets, could be important to support the patterns observed here. The main problem, in this case, is that local sampling is usually more prone to error and more complex methodologies

Table 1. Results of mean pairwise relationship among the 7 sites, according to the methods by Carvalho et al. (2012) and Baselga (2010) for the four groups of terrestrial vertebrates, including mean dissimilarity and proportion in respect to overall β -diversity and results of raw and partial Mantel tests, correlating each pairwise dissimilarity and geographic distances (distance) and biome effect (biome), as well as partial correlations between dissimilarity and geographic distance keeping biome effect constant (geo.biome) and between dissimilarity and biome keeping geographic distances constant (biome.geo).

		Mean	Prop	Distance	Biome	geo.biome	biome.geo
Anf							
<i>Carvalho</i>				r			
Total	Bcc	0.528	1.000	0.761**	0.858**	0.085	0.609**
Richness	Brich	0.266	0.460	0.554*	0.332	0.559*	-0.185
Turnover	B3	0.285	0.540	0.486*	0.486	0.152	0.088
<i>Baselga</i>							
Total	Bsor	0.419	1.000	0.710**	0.841**	-0.051	0.603**
Richness	Bnest	0.117	0.279	0.444	0.472	0.084	0.114
Turnover	Bsim	0.301	0.721	0.739**	0.902**	-0.171	0.914**
Rept							
<i>Carvalho</i>							
Total	Bcc	0.492	1.000	0.840**	0.856**	0.392	0.473*
Richness	Brich	0.117	0.191	-0.233	-0.197	-0.127	0.004
Turnover	B3	0.398	0.809	0.799**	0.804**	0.353	0.325
<i>Baselga</i>							
Total	Bsor	0.371	1.000	0.793**	0.848**	0.233	0.512*
Richness	Bnest	0.057	0.154	-0.318	-0.312	-0.102	-0.042
Turnover	Bsim	0.315	0.846	0.766**	0.824**	0.196	0.452*
Aves							
<i>Carvalho</i>							
Total	Bcc	0.514	1.000	0.613**	0.777**	-0.175	0.501*
Richness	Brich	0.239	0.420	0.306	0.49	-0.261	0.273
Turnover	B3	0.298	0.580	0.624**	0.701**	0.056	0.294
<i>Baselga</i>							
Total	Bsor	0.407	1.000	0.596**	0.766	-0.194	0.490*
Richness	Bnest	0.102	0.251	-0.039	0.073	-0.201	0.107
Turnover	Bsim	0.305	0.749	0.611**	0.758	-0.126	0.449*
Mam							
<i>Carvalho</i>							
Total	Bcc	0.541	1.000	0.650*	0.71	0.108	0.281
Richness	Brich	0.197	0.320	0.170	0.119	0.134	-0.028
Turnover	B3	0.368	0.680	0.578*	0.695	-0.056	0.336
<i>Baselga</i>							
Total	Bsor	0.408	1.000	0.638*	0.705	0.086	0.285
Richness	Bnest	0.097	0.238	-0.318	0.009	-0.641*	0.298
Turnover	Bsim	0.311	0.762	0.655**	0.769	-0.022	0.424*

*P < 0.05; **P < 0.01.

References

- ALMEIDA-NETO, M., FRENSEL, DMB. and ULRICH, W., 2012. Rethinking the relationship between nestedness and beta-diversity: a comment on Baselga (2010). *Global Ecology and Biogeography*, vol. 21, no. 7, p. 772-7. <http://dx.doi.org/10.1111/j.1466-8238.2011.00709.x>
- ALMEIDA-NETO, M., GUIMARÃES, P., GUIMARÃES-JUNIOR, PR., LOYOLA, RD. and ULRICH, W., 2008. A consistent metric for nestedness analysis in ecological systems: reconciling concept and quantification. *Oikos*, vol. 117, p. 1227-1239. <http://dx.doi.org/10.1111/j.0030-1299.2008.16644.x>
- ANDERSON, MJ., CRIST, TO., CHASE, JM., VELLEND, M., INOUE, BD., FREESTONE, AL., SANDERS, NJ., CORNELL, HV., COMITA, LS., DAVIES, KF., HARRISON, SP., KRAFT, NJB., STEGEN, JC. and SWENSON, NG., 2011. Navigating the multiple meanings of b diversity: a roadmap for the practicing ecologist. *Ecology Letters*, vol. 14, p. 19-28. PMID:21070562. <http://dx.doi.org/10.1111/j.1461-0248.2010.01552.x>
- ANDREW, ME., WULDER, MA., COOPS, NC. and BAILLARGEON, G., 2012. Beta-diversity gradients of butterflies along productivity gradients. *Global Ecology and Biogeography*, vol. 21, p. 352-364. <http://dx.doi.org/10.1111/j.1466-8238.2011.00676.x>
- BASELGA, A., 2010. Partitioning the turnover and nestedness components of beta-diversity. *Global Ecology and Biogeography*, vol. 19, p. 134-143. <http://dx.doi.org/10.1111/j.1466-8238.2009.00490.x>
- , 2012. Separating the contribution of species replacement and species richness differences to beta diversity. *Global Ecology and Biogeography*, vol. 21. In press.
- CARVALHO, JC., CARDOSO, P. and GOMES, P., 2012. Determining the relative roles of species replacement and species richness differences in generating beta-diversity patterns. *Global Ecology and Biogeography*, vol. 21, no. 7, p. 760-771. PMID:21352441. <http://dx.doi.org/10.1111/j.1466-8238.2011.00694.x>
- COTTENIE, K., 2005. Integrating environmental and spatial processes in ecological community dynamics. *Ecology Letters*, vol. 8, p. 1175-1182. <http://dx.doi.org/10.1111/j.1461-0248.2005.00820.x>
- DINIZ-FILHO, JAF., SIQUEIRA, T., PADIAL, AA., RANGEL, TF., LANDEIRO, VL. and BINI, LM., 2012. Spatial autocorrelation analysis allows disentangling the balance between neutral and niche processes in metacommunities. *Oikos*, vol. 121, no. 2, p. 201-210. <http://dx.doi.org/10.1111/j.1600-0706.2011.19563.x>
- DOBROVOLSKI, R., MELO, AS., CASSEMIRO, FAS. and DINIZ-FILHO, JAF., 2012. Climatic history and dispersal ability explain the relative importance of turnover and nestedness components of beta-diversity. *Global Ecology and Biogeography*, vol. 21, no. 2, p. 191-7. PMID:17472910 PMCid:2169276. <http://dx.doi.org/10.1111/j.1466-8238.2011.00671.x>
- Eletronorte, 1985a. *Plano de Enchimento do Reservatório: Fauna - Projetos Faunísticos em Tucuruí*. Brasília. vol. 1.
- , 1985b. *Plano de Enchimento do Reservatório: Fauna - Análise da Operação Curupira*. Brasília. vol. 2.
- GASTON, KJ., DAVIES, RG., ORME, CDL., OLSON, VA., THOMAS, GH., DING, TS., RASMUSSEN, PC., LENNON, JJ., BENNETT, PM., OWENS, IPF. and BLACKBURN, TM., 2007. Spatial turnover in the global avifauna. *Proceedings of the Royal Society B: Biological Sciences*, vol. 274, p. 1567-1574. <http://dx.doi.org/10.1098/rspb.2007.0236>
- HARRISON, S., ROSS, SJ. and LAWTON, JH., 1992. Beta diversity on geographic gradients in Britain. *Journal of Animal Ecology*, vol. 61, p. 151-158. <http://dx.doi.org/10.2307/5518>
- HAWKINS, BA., DINIZ-FILHO, JAF., JARAMILLO, CA. and SOELLER, SA., 2006. Post-Eocene climate change, niche conservatism, and the latitudinal diversity gradient of New World birds. *Journal of Biogeography*, vol. 33, p. 770-780. PMID:21645193. <http://dx.doi.org/10.1111/j.1365-2699.2006.01452.x>
- HORTAL, J., DINIZ-FILHO, JAF., BINI, LM., RODRÍGUEZ, MA., BASELGA, A., NOGUÉS-BRAVO, D., RANGEL, TF., HAWKINS, BA. and LOBO, JS., 2011. Ice age climate, evolutionary constraints and diversity patterns in European dung beetles. *Ecology Letters*, vol. 14, p. 741-748. <http://dx.doi.org/10.1111/j.1461-0248.2011.01634.x>
- KOLEFF, P., LENNON, JJ. and GASTON, KJ., 2003. Are there latitudinal gradients in species turnover? *Global Ecology and Biogeography*, vol. 12, p. 483-498. <http://dx.doi.org/10.1046/j.1466-822X.2003.00056.x>
- LEGENDRE, P. and LEGENDRE, L., 1998. *Numerical Ecology*. Amsterdam: Elsevier.
- LENNON, JJ., KOLEFF, P., GREENWOOD, JJD. and GASTON, KJ., 2001. The geographical structure of British bird distributions: diversity, spatial turnover and scale. *Journal of Animal Ecology*, vol. 70, p. 966-979. <http://dx.doi.org/10.1046/j.0021-8790.2001.00563.x>
- MANLY, BFJ., 1985. *Statistics of Natural Selection*. London: Chapman & Hall.
- MELO, AS., RANGEL, TFLV. and DINIZ-FILHO, JAF., 2009. Environmental drivers of beta-diversity patterns in New-World birds and mammals. *Ecography*, vol. 32, p. 226-236. PMID:17355570. <http://dx.doi.org/10.1111/j.1600-0587.2008.05502.x>
- MITTELBACH, GG., SCHEMSKE, DW., CORNELL, HV., ALLEN, AP., BROWN, JM., BUSH, M., HARRISON, SP., HURLBERT, AH., KNOWLTON, N., LESSIOS, HA., McCAIN, CM., McCUNE, AR., McDADDE, LA., McPEEK, MA., NEAR, TJ., PRICE, TD., RICKLEFS, RE., ROY, K., SAX, DF., SCHLUTER, D., SOBEL, JM. and TURELLI, M., 2007. Evolution and the latitudinal diversity gradient: speciation, extinction and biogeography. *Ecology Letters*, vol. 10, p. 315-331. <http://dx.doi.org/10.1111/j.1461-0248.2007.01020.x>
- PASSAMANI, M., 2004. Levantamento de pequenos mamíferos não-voadores da Região Central do Tocantins. *Humanitas*, vol. 4, p. 55-62.
- PAVAN, D. and DIXO, M., 2004. A Herpetofauna da área de influência do reservatório da Usina Hidrelétrica Luís Eduardo Magalhães. *Humanitas*, vol. 4, p. 13-30.
- PINHEIRO, RT., 2004. Captura e recaptura de aves na Região Central do Tocantins. *Humanitas*, vol. 4, p. 39-54.
- PUORTO, G. and BARBARINI, CC., 2004. Coleção herpetológica do Museu de Zoologia do Centro Universitário Luterano de Palmas - MZCEULP. *Humanitas*, vol. 4, p. 31-38.
- QIAN, H., 2009. Beta diversity in relation to dispersal ability for vascular plants in North America. *Global Ecology and Biogeography*, vol. 18, p. 327-332. <http://dx.doi.org/10.1111/j.1466-8238.2009.00450.x>

- RANGEL, TFLVB. and DINIZ-FILHO, JAF., 2005. An evolutionary tolerance model explaining spatial patterns in species richness under environmental gradients and geometric constraints. *Ecography*, vol. 28, p. 253-263. <http://dx.doi.org/10.1111/j.0906-7590.2005.04038.x>
- RANGEL, TFLVB., DINIZ-FILHO, JAF. and BINI, LM., 2006. Towards an integrated computational tool for spatial analysis in macroecology and biogeography. *Global Ecology and Biogeography*, vol. 15, p. 321-327. <http://dx.doi.org/10.1111/j.1466-822X.2006.00237.x>
- RANGEL, TFLVB., DINIZ-FILHO, JAF. and COLWELL, RK., 2007. Species richness and evolutionary niche dynamics: a spatial pattern-oriented simulation experiment. *American Naturalist*, vol. 274, p. 165-174.
- , 2010. SAM: a comprehensive application for spatial analysis in macroecology. *Ecography*, vol. 33, p. 46-50. <http://dx.doi.org/10.1111/j.1600-0587.2009.06299.x>
- SILVA JUNIOR, NJ. and SITES, JW., 1995. Patterns of diversity of neotropical squamate reptile species with emphasis on the Brazilian Amazon and the conservation potential of indigenous reserves. *Conservation Biology*, vol. 9, p. 873-901. <http://dx.doi.org/10.1046/j.1523-1739.1995.09040873.x>
- SOININEN, J., McDONALD, R. and HILLEBRAND, H., 2007. The distance decay of similarity in ecological communities. *Ecography*, vol. 30, p. 3-12.
- Systema Naturae Consultoria Ambiental Ltda., 1996. *Inventariamento da Fauna Silvestre*. Usina Hidrelétrica Serra da Mesa. Goiânia. Relatório Final.
- , 1999. *Operação Lobo Guará - Resgate da Fauna*. Usina Hidrelétrica Serra da Mesa. *Relatório Final*. Goiânia, GO, Brasil.
- , 2008. *Monitoramento de Animais Silvestres - Etapas Pré e Pós-enchimento do Reservatório*. Usina Hidrelétrica Peixe Angical. Goiânia. Relatório Final.
- , 2010a. *Programa da Fauna Silvestre - Monitoramento Faunístico Pós-enchimento - Fase IV - Ano VII*. Usina Hidrelétrica Cana Brava. Goiânia. Relatório Final.
- , 2010b. *Programa de Monitoramento e Conservação da Fauna*. Usina Hidrelétrica São Salvador. Goiânia. Relatório Final.
- , 2011. *Programa de Monitoramento da Fauna: Complementação do Levantamento Faunístico nas Áreas dos Ribeirões Mosquito, João Aires, Curicaca, Córrego Jatobá e Rio Feio*. Usina Hidrelétrica Estreito. Goiânia. Relatório Técnico Interpretativo 2.
- TUOMISTO, H., 2010a. A diversity of beta diversities: straightening up a concept gone awry. Part 1. Defining beta diversity as a function of alpha and gamma diversity. *Ecography*, vol. 33, p. 2-22. <http://dx.doi.org/10.1111/j.1600-0587.2009.05880.x>
- , 2010b. A diversity of beta diversities: straightening up a concept gone awry. Part 2. Quantifying beta diversity and related phenomena. *Ecography*, vol. 33, p. 23-45. <http://dx.doi.org/10.1111/j.1600-0587.2009.06148.x>
- VASCONCELOS, TS., RODRIGUEZ, MA. and HAWKINS, BA., 2011. Biogeographic distribution patterns of South American amphibians: a regionalization based on cluster analysis. *Natureza & Conservação*, vol. 9, p. 67-72.
- WHITTAKER, RH., 1972. Evolution and measurement of species diversity. *Taxon*, vol. 21, p. 213-251. <http://dx.doi.org/10.2307/1218190>
- WIENS, JJ. and DONOGHUE, MJ., 2004. Historical biogeography, ecology and species richness. *Trends in Ecology & Evolution*, vol. 19, p. 639-644. <http://dx.doi.org/10.1016/j.tree.2004.09.011>